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## Land use Planning in Ireland: a Life Cycle Energy Analysis of Recent Residential Development in the Greater Dublin Area

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# Land use planning in Ireland—a life cycle energy analysis of recent residential development in the Greater Dublin Area

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## Abstract

**Background, aim, and scope** One third of the total housing stock in the Republic of Ireland has been built in 10 years up to and including 2006 and of this approximately 34% was built in the Greater Dublin Area (GDA). Much of the housing was low-density with poor public transport links leading to doubts over its sustainability—particularly in terms of energy use. Although the country is committed to reducing greenhouse gases to 13% above 1990 levels by the period 2008–2012, by 2005, emissions were already 25.4% higher than the baseline and current projections are that this figure will rise to 37% over the period. The residential sector is estimated to contribute to approximately 24.5% of energy-related CO<sub>2</sub> emissions. This paper estimates total emissions from residential developments in the GDA constructed between 1997 and 2006.

**Materials and methods** Carbon dioxide equivalent (CO<sub>2</sub>) emissions are estimated using a life cycle assessment approach over a 100-year building lifespan and employing process, input–output and hybrid energy techniques. Life cycle stages include: construction, operation, transport, maintenance and demolition. The main data sources include: national population and industry census data, household travel survey data, residential energy performance surveys and national accounts. The GDA was split into four zones each encompassing development at increas-

ing radii from Dublin's city centre, namely: city centre, suburbs, exurbs and commuter towns.

**Results** Per capita CO<sub>2</sub> life cycle emissions in the GDA were found to be approximately 50–55% greater in the exurbs and commuter towns than in the city centre. Of the five life cycle stages studied, operational energy requirements (predominantly space heating and hot water, but including power) contributed most significantly to emissions (68%), followed by transport (17%), construction (9%) and maintenance/renovation (6%).

**Discussion** Operating emissions from dwellings in the commuter town and extra-urban zones were almost twice those in the city centre both due to larger dwelling sizes and the predominance of detached and semi-detached dwellings (with large amounts of exposed walls) in the former and the prevalence of smaller apartments in the latter. Car use was most pronounced in the zones furthest from the city centre where per capita emissions were almost twice those of residents in the city centre. Despite their smaller size, the per capita construction CO<sub>2</sub> emissions of apartments were approximately one third greater than for low-rise dwellings due to the greater energy intensity of the structure. However, this difference was more than compensated for by the significantly lower operational emissions referred to above.

**Conclusions** In 2006, recurrent CO<sub>2</sub> emissions (operational, transport and maintenance) from dwellings built in the GDA over the ten preceding years were 2,108 kt while construction-related emissions in that year were 1,325 kt giving a total contribution from the residential sector of 3,434 kt CO<sub>2</sub>/annum—representing 4.9% of national emissions for that year. Had the development policy prescribed 'city centre'-type development and transport modes, then emissions for the year 2006 would have been 2,892 kt CO<sub>2</sub>—a reduction of almost 16% over the actual figure. However, in this scenario recurrent emissions would have

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71 been reduced to 1,417 kt CO<sub>2</sub>—a reduction of 33% over  
 72 actual levels.  
 73 *Recommendations and perspectives* This study supports  
 74 Irish and international governments’ policies aimed at  
 75 curbing CO<sub>2</sub> emissions from the domestic sector which  
 76 focus primarily on reducing operational emissions from  
 77 new and existing housing through design and construction  
 78 improvements. However, it demonstrates that significant  
 79 reductions in operational emissions are associated with  
 80 high-density residential development with modest floor  
 81 areas. Furthermore, it highlights the scope for transport  
 82 emissions’ reductions through better spatial planning  
 83 leading to reduced car travel.

84 **Keywords** Carbon dioxide equivalent · CO<sub>2</sub> · Domestic  
 85 dwellings · Embodied energy · Energy · Greenhouse gas  
 86 emissions · Life cycle assessment · Spatial planning

87 **1 Background, aim, and scope**

88 At the end of 2006 there were 1,835,515 domestic dwell-  
 89 ings in the Republic of Ireland (CSO 2007a). Of these,  
 90 607,961—representing one third of the total housing  
 91 stock—were built in the 10 years up to and including that  
 92 year (Fig. 1). Approximately 34% of this new housing was  
 93 built in Dublin and the greater Dublin region, much of  
 94 which was low-density with poor public transport links.  
 95 Indeed, much of the housing development in Ireland can be  
 96 characterised as ‘once-off’ or low-density located outside  
 97 urban centres: over one third of the development comprised  
 98 detached houses, 44% were semi-detached or terraced  
 99 housing units and only one fifth were apartments. In  
 100 contrast, multi-family dwellings account for almost half of  
 101 Europe’s housing stock (Netherlands Ministry of Housing,  
 102 Spatial Planning and the Environment 2004). The adoption

in Ireland of such low density suburban and extra-urban  
 development policies may have resulted in increased  
 greenhouse gas emissions when compared with higher  
 density urban developments.

Ireland’s energy emissions’ performance is poor by  
 international standards: in 2004, energy use per capita was  
 3,870 kg of oil equivalent (kgoe) and associated greenhouse  
 gas emissions were 10,589 kg. This compares unfavourably  
 with EU27 figures of 3,689 kgoe and 8,180 kg, respectively  
 (Eurostat 2007). Although the country has committed to  
 reducing greenhouse gases to 13% above 1990 levels by  
 the period 2008–2012, by 2005, emissions were already  
 25.4% higher than the baseline (EPA 2007) and current  
 projections are that this figure will rise to 37% over the  
 period. Residential energy use accounted for almost one  
 quarter of total national fuel consumption in 2005 of which  
 almost 75% is used for space heating and hot water  
 (Howley et al. 2006).

The impact of a domestic development on greenhouse  
 gas emissions is related to the quantity and carbon content  
 of the energy required over its lifetime. For example, two  
 similar buildings with different passive thermal perform-  
 ances will consume different amounts of energy: one will  
 emit more greenhouse gases than the other. Identical  
 buildings powered by fuels with different carbon contents  
 (for example, biomass- versus oil-fuelled boilers) will also  
 have differing impacts. The energy impact of a domestic  
 development over its entire life cycle can be viewed as the  
 sum of this operational energy use together with the  
 energy needed to produce and maintain it, demolish it as  
 well as the energy required to travel to and from it.  
 Similarly, the CO<sub>2</sub> impact is the global warming potential-  
 weighted sum of greenhouses gases (predominantly  
 carbon dioxide, nitrous oxide and methane) emitted by  
 this energy use.

Australian literature suggests that the average energy  
 required to produce a house in that country is approximately  
 5 GJ/m<sup>2</sup> producing some 0.49 tonnes/m<sup>2</sup> floor area of carbon  
 dioxide (CSIRO 2007) based on Australian emissions’  
 intensities. This is referred to as ‘embodied energy’ and  
 theoretically includes all direct and indirect production  
 energy inputs such as construction, material production,  
 raw material extraction and associated services. The  
 average floor area of a domestic dwelling in Ireland in  
 2004 was 112 m<sup>2</sup> (SEI 2005) and applying the Australian  
 data, this gives a per-dwelling embodied energy of  
 560 GJ and emissions of 55 tonnes of carbon dioxide  
 (CO<sub>2</sub>). Assuming a 100-year building lifespan, this is  
 equal to annualised emissions of 0.55 tonnes of CO<sub>2</sub> and  
 5.6 GJ of energy use. Although these figures are rough  
 estimates—since the Australian embodied energies are  
 likely to differ from those in Ireland (no Irish data are  
 available at the time of writing)—when used with the

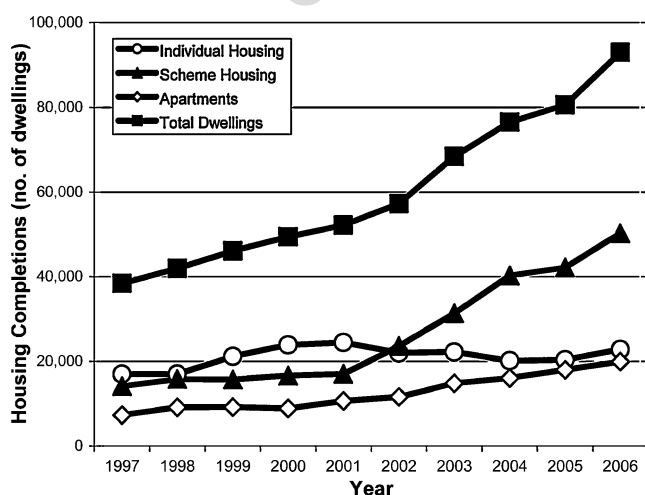


Fig. 1 Annual housing completions in Ireland 1997–2006

156 operational and travel data below, they indicate the  
 157 relative importance of embodied energy in the life cycle  
 158 of a domestic building and will assist in the development  
 159 of the life cycle assessment (LCA) methodology described  
 160 in this paper.  
 161 In 2004, the mean energy use per Irish dwelling was  
 162 88 GJ resulting in the emission of 8.2 tonnes of CO<sub>2</sub> per  
 163 annum. Some 79% of energy use was in the form of direct  
 164 fossil fuel consumption, the remainder was electricity (SEI  
 165 2005).  
 166 In 2006, the Irish transport sector consumed 213,000 TJ  
 167 of energy and emitted 15,273 kt of CO<sub>2</sub> (SEI 2006) of  
 168 which 40% was attributable to private cars (SEI 2004). In  
 169 the same year, there were 1,469,521 households in Ireland  
 170 (CSO 2007b) giving average annual transport-related CO<sub>2</sub>  
 171 and energy use figures of 4.2 tonnes and 58 GJ per  
 172 dwelling, respectively. This approximation overestimates  
 173 the direct energy inputs associated with moving to and  
 174 from a domestic dwelling since many private car trips  
 175 would be used for other purposes, although it ignores  
 176 indirect energy inputs such as product development and  
 177 manufacturing.  
 178 Little data are available on the maintenance-related  
 179 energy requirements of domestic dwellings and associated  
 180 emissions. Figures for the refurbishment and repair of office  
 181 buildings are reported to lie in the range of 0.17 to 0.34 GJ/  
 182 m<sup>2</sup> (Yohanis and Norton 2002). Taking the lower figure—  
 183 since office refurbishment is likely to be more energy-  
 184 intensive than domestic refurbishment—and assuming an  
 185 interval between refurbishments of 7 years, the annual  
 186 maintenance-related energy intensity for an average Irish  
 187 house is 2.72 GJ/m<sup>2</sup> with associated emissions of 0.23 t  
 188 CO<sub>2</sub>. However, even the use of the lower office refurbish-  
 189 ment energy intensity figure may overestimate domestic  
 190 emissions.  
 191 Based on these data, the total approximate annualised  
 192 life cycle energy requirement for an average domestic Irish  
 193 dwelling is 154.3 GJ emitting 13.2 tonnes of CO<sub>2</sub>.  
 194 Transport and operation dominate life cycle energy require-  
 195 ments (38% and 56% respectively) and CO<sub>2</sub> emissions  
 196 (32% and 62% respectively). Based on these preliminary  
 197 results, the methodology outlined in this paper focuses  
 198 therefore to a greater extent on operational and transport  
 199 energy use and emissions than on construction, mainte-  
 200 nance and demolition.  
 201 The objectives of this study are to: determine the whole-  
 202 life energy and CO<sub>2</sub> intensities of dwellings built in the  
 203 Greater Dublin Area (GDA) between 1997 and 2006; assess  
 204 the relative impact of different life cycle stages on energy  
 205 use and CO<sub>2</sub> emissions; and quantify and compare the  
 206 impact of urban, suburban and extra-urban residential  
 207 developments in and around Dublin on national energy-  
 208 related greenhouse gas emissions.

**2 Materials and methods** 209

Life cycle assessment (LCA) was used to determine energy 210  
 use and carbon dioxide emissions for all stages of a 211  
 residential development including: 212

- construction; 213
- operation including heating, hot water, lighting and 214  
 small power loads; 215
- travel to and from the development; 216
- maintenance; and 217
- demolition 218

The study only considered the greenhouse gas impacts 219  
 associated with energy use and ignored wider environmen- 220  
 tal impacts such as resource depletion, groundwater 221  
 pollution and habitat loss. CO<sub>2</sub> emissions only were 222  
 calculated since these data are easily available. However, 223  
 since CO<sub>2</sub> accounts for only 93% of non-agricultural 224  
 greenhouse gas (GHG) emissions in Ireland by global 225  
 warming potential (CSO, 2007c), resulting figures were 226  
 divided by 0.93 to correct for unaccounted energy-related 227  
 GHG emissions (such as N<sub>2</sub>O and CH<sub>4</sub>) to estimate total 228  
 emissions in CO<sub>2</sub> equivalent (referred to hereafter simply as 229  
 CO<sub>2</sub>). 230

Energy requirements and emissions for each of the life 231  
 stages of the residential development were determined 232  
 using process, input–output or hybrid process analysis 233  
 depending on data availability. These methodologies are 234  
 described by inter alia Crawford (2007) and Bullard et al. 235  
 (1978). 236

**2.1 Developments** 237

Two developments were selected to represent typical recent 238  
 residential developments in the Greater Dublin Area (a term 239  
 which is used to describe the city and county of Dublin as 240  
 well as the adjacent counties of Kildare, Meath and 241  
 Wicklow) which are described below. 242

A high density mixed-use urban development in the city 243  
 of Dublin comprising 300 apartments of one, two and three 244  
 bedrooms with a total gross floor area of 22,500 m<sup>2</sup>. The 245  
 scheme also contains retail and office units. 246

A development of 118 two-storey detached, semi- 247  
 detached and terraced houses on a 5.7 ha site outside the 248  
 city comprising detached, semi-detached and terraced 249  
 houses ranging in size from three to five bedrooms. 250

**2.2 Construction** 251

The energy embodied in the above developments was 252  
 determined using hybrid process analysis up to the point of 253  
 completion of the construction process by the contractor 254  
 and hand over to the new resident. This required material- 255



256 specific energy intensities (Hammond and Jones 2006)  
 257 which were used in conjunction with national input–output  
 258 tables (CSO 2006a) to quantify the total upstream and  
 259 direct energy requirements for the building. Upstream  
 260 energy requirements might include raw materials’ extrac-  
 261 tion and building materials’ manufacturing; direct energy  
 262 relates to energy expended during the construction process  
 263 itself.

264 Bills of quantities detailing material quantities and prices  
 265 were obtained from the contractor who built the schemes.  
 266 Where possible, materials were identified, characterised and  
 267 process energy intensities applied. Energy intensities of  
 268 unidentified expenditure were estimated using Irish con-  
 269 struction sector energy intensities derived from national  
 270 input–output tables, mean national energy tariffs, primary  
 271 energy factors and disaggregation coefficients (Acquaye  
 272 et al. 2008). An integral part of this methodology involved  
 273 calculating a carbon dioxide equivalent intensity based on  
 274 the mix of fuels used.

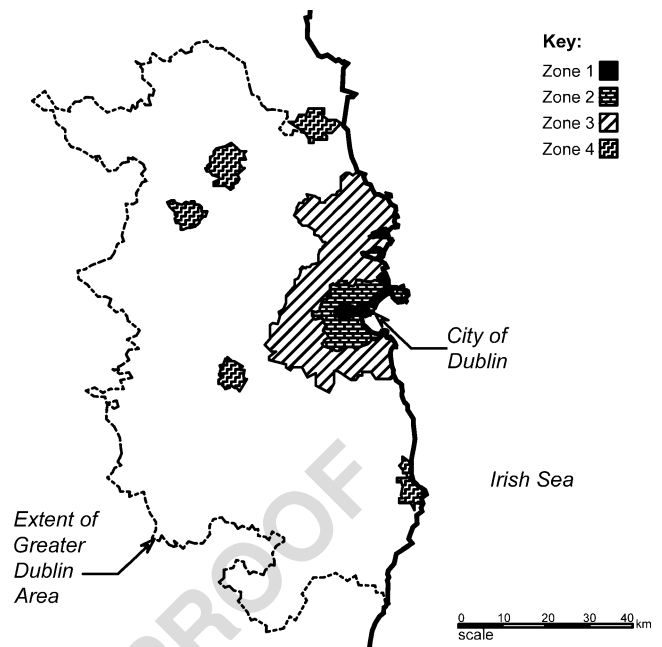
275 **2.3 Transport**

276 Transport emissions for all trips to and from dwellings were  
 277 calculated based on data from the Dublin Transportation  
 278 Office’s 2006 Household Survey (DTO 2006) of approxi-  
 279 mately 2,500 households in the Greater Dublin Area. The  
 280 survey was carried out over a 1-week period where  
 281 participants kept a detailed log of all movements to and  
 282 from their dwellings. Relevant variables were extracted  
 283 from the database which included:

- 284 • mode of transport (for example train, rapid rail, tram,
- 285 bus, truck, van, car, taxi, motorbike, running, walking);
- 286 • number of trips made;
- 287 • trip distances; and
- 288 • trip duration

289 It is accepted that development density and land-use mix  
 290 are inversely related to household-related travel distances  
 291 and vehicle emissions (Frank et al. 2000; Næss 2006;  
 292 Newman and Kenworthy 1999). Moreover, the distance of  
 293 a residence from concentrated employment and recreational  
 294 areas (typically a city centre) affects both transport mode  
 295 (DTO 2006) and travel distance and, therefore, fuel  
 296 consumption and CO<sub>2</sub> emissions. In order to capture the  
 297 effects of density and distance to urban centres, the Greater  
 298 Dublin Area Household Survey data were therefore divided  
 299 into four zones of increasing distance from the city centre  
 300 (Fig. 2). These are:

302 **Zone 1** City Centre which is roughly bounded by two  
 303 canals and the inner ring roads to the north and  
 304 south giving a radius of up to 3.0 km around the  
 305 city;



**Fig. 2** Map of the Greater Dublin Area (GDA) showing the zones analysed

- 306 **Zone 2** Suburbs which are located between the City 306
- 307 Centre and the M50 motorway ring road at a 307
- 308 radius of approximately 9.0 km around the city; 308
- 309 **Zone 3** Exurbs covering the low density urban areas in the 309
- 310 remainder of county Dublin located outside the 310
- 311 M50 incorporating many of the closer commuter 311
- 312 towns at a radius of 15 to 30 km from the centre; 312
- 313 and 313
- 314 **Zone 4** Commuter towns of Drogheda (42 km from the 314
- 315 centre), Navan (45 km), Enfield (40 km), Naas 315
- 316 (31 km) and Wicklow (43 km). 316

317  
 318 Data were then analysed by zone to determine the total 318  
 319 number of trips by mode of transport and total annualised 319  
 320 modal distances travelled. Mean data were then calculated 320  
 321 on a per household and per capita basis. 321

322 Fuel consumption and CO<sub>2</sub> emissions data were calcu- 322  
 323 lated in the following ways: 323

324 no carbon dioxide emissions or fuel consumption was 324  
 325 attributed to walking, running or cycling. This assumes 325  
 326 that the energy and CO<sub>2</sub> embodied in equipment such 326  
 327 as footwear, bicycles, special clothing and helmets is 327  
 328 negligible compared to that embodied in other forms of 328  
 329 transport. The food requirements of individuals are 329  
 330 outside the LCA boundary in all cases. 330  
 331 aeroplane, train, rapid rail transit, tram, bus, mini-bus, 331  
 332 truck and motorbike energy uses and emissions were 332  
 333 calculated using published average emissions per 333  
 334 person kilometre (Cox and Hickman 1998; Davies 334

335 and Diegel 2007). Trains, buses, mini-buses and trucks  
 336 were assumed to be diesel-fuelled while DART (a rapid  
 337 rail system) and LUAS (a tram system) use electricity  
 338 only. The average fuel consumption and carbon  
 339 dioxide emissions data used in the analysis is presented  
 340 in Table 1.

341 Since emissions from cars were expected to predomi-  
 342 nate, these were calculated in the more detailed manner  
 343 described below.

344 Engine size and fuel type (diesel/petrol) distributions  
 345 were determined using 2005 local government car registra-  
 346 tion data and a weighted mean engine size was calculated  
 347 for both fuel types.

348 Mean fuel consumption was calculated for this engine  
 349 size based on published data and speed-related efficiencies  
 350 (Davies and Diegel 2007) were combined to give an  
 351 equation for fuel consumptions for the weighted mean  
 352 engine size (calculated above) at different speeds.

353 Average speeds for all car trips were calculated and fuel  
 354 consumption determined using the fuel consumption to  
 355 speed relationship.

356 This represents the fuel consumed only and does not  
 357 consider the full life-cycle energy use of the car. Research  
 358 by Castro et al. (2003) suggests that direct fuel consump-  
 359 tion represents approximately 95% of the total life cycle  
 360 energy requirements of a car and the results were adjusted  
 361 accordingly. In the absence of equivalent life cycle  
 362 assessment literature for other modes of transport, this  
 363 figure was also applied to buses, trains, vans, trucks and  
 364 motorbikes.

365 **2.4 Operation**

366 Operational energy requirements were quantified using the  
 367 Energy Performance Survey of Irish Housing undertaken  
 368 jointly by Sustainable Energy Ireland and CODEMA  
 369 between 2004 and 2005. This study involved a detailed  
 370 survey of the energy use characteristics of dwellings in  
 371 Ireland and covered a range of construction types, ages and  
 372 occupancy patterns. Relevant survey data were selected to  
 373 be representative of the development types chosen here, in  
 374 particular:

375 Data in the fields ‘Detached’, ‘Semi-Detached’,  
 376 ‘Terraced’ and ‘Purpose-Built Apartment’ were incor-

porated whereas, ‘Converted Apartment’ and ‘Other’ 377  
 were ignored and 378  
 Developments built since 1997 were included since 379  
 different Building Regulations and energy performance 380  
 standards pertained prior to this date and would not 381  
 have been representative of the 1997–2006 housing 382  
 stock. However, a lack of data for apartments neces- 383  
 sitated the use of pre-1997 data which would have 384  
 overestimated energy use and emissions to some 385  
 extent. 386

Mean annualised energy use and carbon dioxide emis- 387  
 sions data were calculated for the reduced sample of 59 388  
 dwellings based on recorded fuel mixes and quantities and 389  
 these were validated against Scottish data. Data were 390  
 converted from per-dwelling to per capita by dividing by 391  
 the average number of persons per private household for 392  
 the county or city as listed in the 2006 national census 393  
 (CSO 2007a). 394  
 395

**2.5 Maintenance and demolition** 396

Maintenance requirements were determined using input– 397  
 output analysis. Total expenditure for Irish housing repair, 398  
 maintenance and improvements for the year 2004 were 399  
 used (CSO 2006b) together with the energy intensity for the 400  
 construction sector and total existing housing units for the 401  
 period to determine the average annual domestic dwelling 402  
 maintenance energy use. 403

Demolition costs were calculated using input–output 404  
 analysis employing data from the 2004 Census of Building 405  
 Construction (CSO 2006c). Turnover and input figures for 406  
 Nace 45.1 ‘Site preparation, demolition and wrecking of 407  
 buildings, earth moving, test drilling and boring’ were used 408  
 to determine energy intensities and these were used with 409  
 2004 published demolition costs (Davis Langdon and 410  
 Everest 2004) to determine demolition energy requirements. 411

**3 Results and discussion** 412

**3.1 Construction** 413

Figure 3 shows the embodied CO<sub>2</sub> emissions per dwelling. 414  
 It can be seen that detached dwellings were found to have 415

t1.1 **Table 1** Average fuel consump-  
 tions and CO<sub>2</sub> emissions for  
 various modes of transport ex-  
 cluding cars (source: EEA 2003)

Mode	Fuel consumption (grams oil equivalent/passenger km)	CO <sub>2</sub> emissions (g/passenger km)	
Bus	23	73	t1.3
Truck	97	216	t1.4
Motorbike	36	104	t1.5
Rail	20	44	t1.6

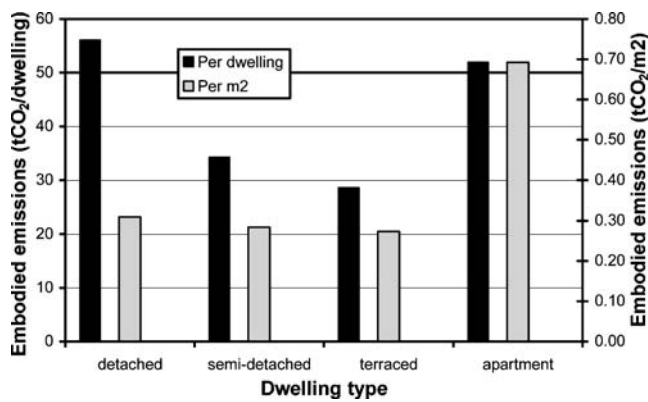


Fig. 3 Embodied CO<sub>2</sub> for each dwelling type studied expressed both on a per dwelling and a per m<sup>2</sup> of gross floor area basis

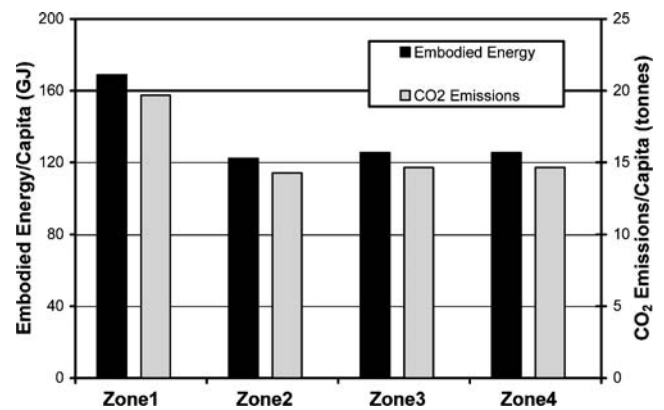


Fig. 5 Per capita embodied energy and construction-related CO<sub>2</sub> emissions for each zone

416 the highest embodied emissions due to their large size and  
 417 construction characteristics. For example, the mean size of  
 418 a detached house in the sample was 181.2 m<sup>2</sup> whereas  
 419 semi-detached houses, terraced houses and apartments were  
 420 120.6, 104.6 and 75.0 m<sup>2</sup>, respectively. Detached houses  
 421 have greater construction materials' requirements since they  
 422 have no shared walls, unlike other dwelling types. Apart-  
 423 ments have the second highest embodied energy due to the  
 424 need for more structural elements and, consequently,  
 425 greater quantities of steel and concrete—both materials  
 426 which have relatively high energy intensities. The use of  
 427 these materials results in apartments producing the most  
 428 CO<sub>2</sub> emissions per constructed dwelling. Terraced houses  
 429 were found to have the lowest embodied energy and CO<sub>2</sub>  
 430 emissions due to their relatively small size and use of  
 431 shared structure (such as party walls). When compared on a  
 432 per square metre of constructed gross floor area, it can be  
 433 seen that due to their relatively small size and large  
 434 structural requirements, apartments have almost two and a  
 435 half times more embodied energy and CO<sub>2</sub> emissions than  
 436 other forms of residential construction.

437 Figure 4 shows the percentage of house type completed  
 438 in 2006 by zone where it can be seen that areas closest to

439 urban centres had the highest concentration of apartment  
 440 completions: almost four fifths of dwellings completed in  
 441 the city centre were apartments while this figure fell to 16%  
 442 in Zones 3 and 4 where the development of 'scheme  
 443 houses' comprising mixed housing units dominated.

444 Based on the mix of dwellings completed in the 10-year  
 445 period between 1997 and 2006 and on the average number  
 446 of persons per household (CSO 2007a), average embodied  
 447 energies and CO<sub>2</sub> emissions were calculated for each zone  
 448 and are shown in Fig. 5. It can be seen that both embodied  
 449 energy requirements and CO<sub>2</sub> emissions are highest in Zone  
 450 1 due to its high proportion of energy-intensive apartments  
 451 and relatively low occupancy figures (2.5 persons per  
 452 household compared to 3.0 for Zone 4).

### 3.2 Transport

454 It was found that the further a development is from the city  
 455 centre, the greater the reliance on the private car while less  
 456 trips are made using public transport. In Zone 1 some 31%  
 457 of trips are made by car compared to 49%, 58% and 63%  
 458 for Zones 2, 3 and 4, respectively. The percentage of trips  
 459 made by walking or cycling decreases with increasing  
 460 distance from the city centre: for Zones 1, 2 and 3 the

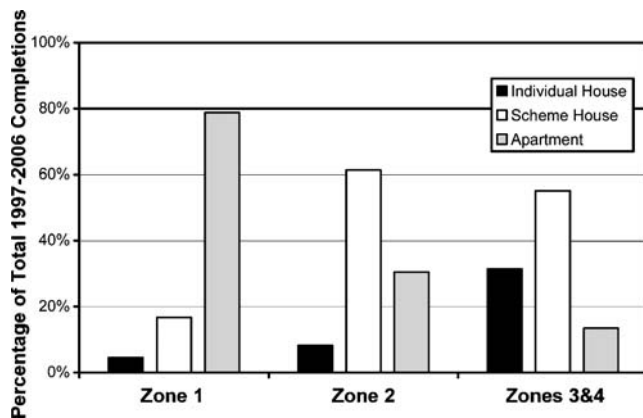


Fig. 4 Percentage of housing completions by type for each zone

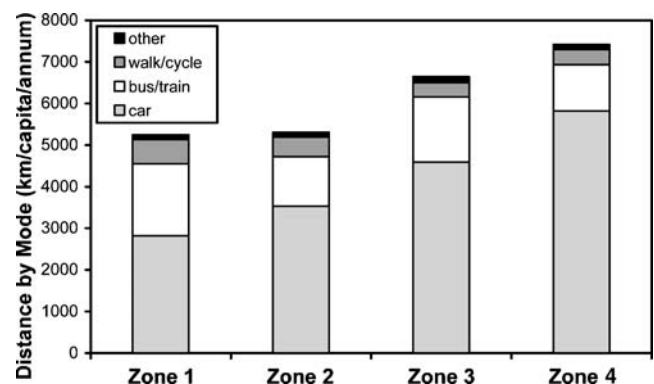
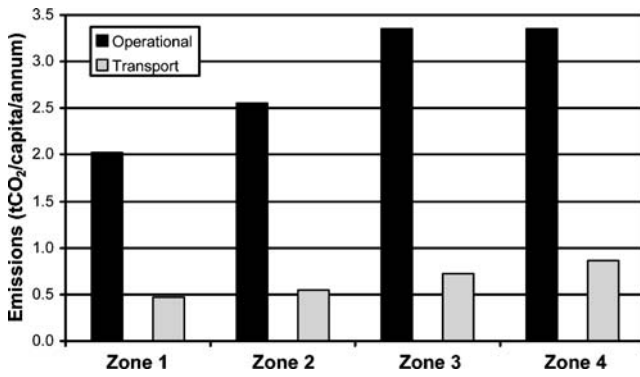


Fig. 6 Annual per capita distance travelled by mode for each zone



**Fig. 7** Annual per-capita operational and transport-related CO<sub>2</sub> emissions for each zone

percentages of trips made using these modes are 40%, 29% and 22%, respectively. However, for Zone 4—commuter towns—the figure increases to 28% suggesting that local employment centres and amenities situated close to residential areas create greater opportunities for walking or cycling than in the Exurbs (Zone 3).

Figure 6 shows the annual modal distances travelled per capita for each zone. There is a clear positive relationship between distance from the centre of Dublin and distance travelled by car. The per capita distances travelled by public transport (bus and train) were highest for Zones 1 and 3 at 1,735 km and 1,566 km, respectively and lowest for Zones 2 and 4 at 1,196 km and 1,119 km, respectively.

Not surprisingly, transport-related fuel consumption and CO<sub>2</sub> emissions also increase the further a residential development is located from the city centre. Figure 7 shows the annual transport-related CO<sub>2</sub> emissions per capita for each zone. It can be seen that annual transport emissions in Zone 4 were highest where the average person consumed 14.2 GJ of energy (equivalent to 449 l of petrol) and emitted 866 kg of CO<sub>2</sub> in meeting their 2006 transport needs. These figures decrease with increased proximity to the city centre and reach a minimum of 8.3 GJ and 472 kg CO<sub>2</sub> for Zone 1. Transport-related CO<sub>2</sub> emissions are therefore almost 85% higher for residential developments in the commuter towns than in the city centre and approximately 16% and 54% higher for suburban (Zone 2) and extra-urban (Zone 3) areas, respectively.

**3.3 Operation** 489

Results for the annual operational (heating, hot water and small power) energy requirements and associated CO<sub>2</sub> emissions for four different dwelling types are shown in Table 2 both on a per square metre and per dwelling basis. Detached houses were found to emit 88 kg CO<sub>2</sub>/m<sup>2</sup>/annum of floor area; the corresponding figures for semi-detached houses, terraced houses and apartments were 67, 55 and 48 kg/m<sup>2</sup>, respectively. Detached houses consume the greatest amounts of energy due to their higher floor areas and completely exposed external envelopes. Conversely, apartments have the lowest per dwelling consumption due to both their smaller size and reduced area of external envelope. Results were compared to emissions statistics for Scotland due to its similar weather and demographics; these are presented in Fig. 8. It can be seen that Scottish emissions compare favourably but are lower for detached and, to a lesser extent, semi-detached houses and are similar for terraced houses and apartments. These differences may be accounted for by the lower CO<sub>2</sub>-intensity of electricity production in Scotland and the relatively large sizes of detached and semi-detached houses constructed in the GDA over the period being analysed.

Figure 7 shows annual operational CO<sub>2</sub> emissions per capita in each zone, taking account of the mix of house types and average occupancies for each zone. On average, each person in Zones 3 and 4 consumes approximately 85% more energy than in Zone 1 and 30% more than in Zone 2 in order to heat and power their home. This is due to the dominance of detached houses in the exurbs and commuter towns and the greater concentration of apartments in the city centre and suburbs. This effect of housing type distribution on increasing per capita operational energy use in Zones 3 and 4 is offset to some degree by higher dwelling occupancy rates in these areas (3.0 per dwelling versus 2.5 for urban).

**3.4 Maintenance and demolition** 525

The average energy requirements associated with maintaining and improving a domestic dwelling were estimated

**Table 2** Annual operational energy requirements and associated CO<sub>2</sub> emissions for different dwelling types

Dwelling type	Energy requirements		CO <sub>2</sub> output		Average size (m <sup>2</sup> )
	(GJ/dwelling/year)	(MJ/m <sup>2</sup> /year)	(t/dwelling/year)	(kg/m <sup>2</sup> /year)	
Apartments	32	400	3.8	48	75
Terraced	65	623	5.8	55	105
Semi-Detached	92	763	8.1	67	121
Detached	161	886	15.9	88	181



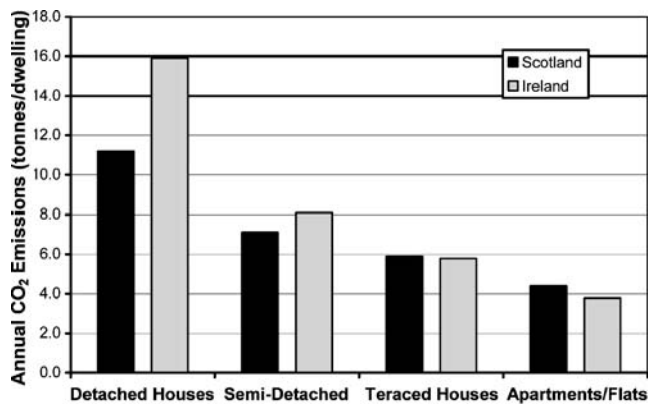


Fig. 8 CO<sub>2</sub> emissions per dwelling for Scottish housing stock and GDA sample (source: Hinchliffe 2005)

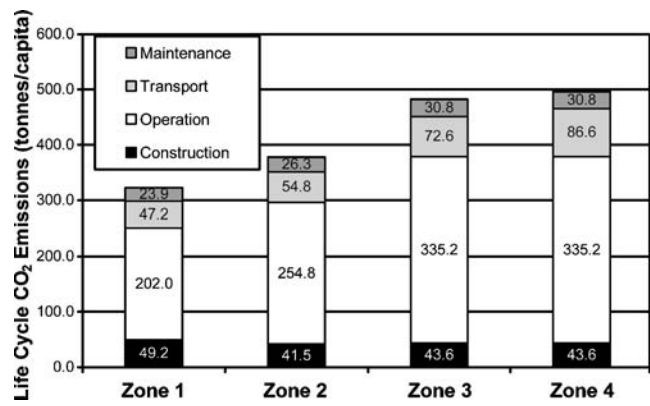


Fig. 9 Life cycle component CO<sub>2</sub> emissions for each zone (demolition has been excluded since emissions are negligible compared to other life cycle stages)

528 using input–output analysis (as described in Section 2.5) to  
 529 be 0.046 GJ/m<sup>2</sup>/annum producing 6.8 kg CO<sub>2</sub>/m<sup>2</sup>/annum.  
 530 Due to a lack of data it was not possible to differentiate  
 531 between maintenance requirements for different dwelling  
 532 types. It is likely that this aggregation problem under-  
 533 estimates maintenance energy requirements for detached  
 534 and semi-detached houses given the greater opportunities  
 535 for the major refurbishment and extension of such  
 536 properties over apartments and—to a lesser extent—  
 537 terraced houses. Therefore, in the absence of more  
 538 disaggregated data, the same per square metre energy and  
 539 CO<sub>2</sub> emissions intensities were applied to the average  
 540 house size for each zone.

541 The cost of demolishing a typical 120 m<sup>2</sup> domestic  
 542 dwelling (brick, block and timber floor construction) was  
 543 estimated at € 17,500 in 2004 and the energy intensity for  
 544 Nace 45.1 was calculated to be 2.25 MJ/€ of demolition  
 545 expenditure. Combining these figures gives a housing  
 546 demolition energy requirement of approximately 0.33 GJ/m<sup>2</sup>  
 547 with associated emissions of 8.3 kg CO<sub>2</sub>/m<sup>2</sup>. The additional  
 548 cost of demolishing the concrete or steel structure in an  
 549 apartment is estimated to be approximately € 14,300,  
 550 resulting in an apartment demolition energy intensity of  
 551 0.60 GJ/m<sup>2</sup> with associated emissions of 15.1 kg CO<sub>2</sub>/m<sup>2</sup>.  
 552 These figures do not allow for recycling.

### 553 3.5 Life cycle

554 Figure 9 shows the CO<sub>2</sub> emissions for a residence built in  
 555 the last 10 years in each of the four zones in the GDA. This  
 556 assumes a building lifespan of 100 years. On average Zone  
 557 3 and 4 occupants consumed 92–98% more energy than  
 558 those in Zone 1 and approximately one third more than  
 559 Zone 2. Per capita CO<sub>2</sub> emissions in Zones 3 and 4 were  
 560 approximately 50–55% greater than Zone 1 and almost  
 561 33% greater than Zone 2. The higher CO<sub>2</sub>/energy ratio in  
 562 Zone 1 can be accounted for by the high incidence of CO<sub>2</sub>-  
 563 intensive electrical storage heating in Zone 1 apartments.

564 Figure 10 shows CO<sub>2</sub> emissions for each dwelling type  
 565 excluding transport but including the construction,  
 566 operation and maintenance life cycle stages (demolition  
 567 has been excluded since these emissions are negligible).  
 568 It can be seen that operational emissions dominate and  
 569 that total emissions over 100 years are significantly  
 570 higher for detached houses (1,773 t) than for semi-  
 571 detached (929 t), terraced (680 t) and apartments (482 t).  
 572 This is due to the greater electrical and space heating  
 573 requirement associated with the large floor areas of  
 574 detached houses in Ireland and their relatively greater  
 575 exposed-envelope/volume ratios. These figures demon-  
 576 strate significant benefits in promoting high-density  
 577 spatial planning policies.

578 When the data for all zones are aggregated, operational  
 579 requirements dominate accounting for 68% of all CO<sub>2</sub>  
 580 emissions followed by transport at 17%, construction at 9%  
 581 and maintenance at 6%. Demolition contributes to less than  
 582 0.5% of life cycle emissions.

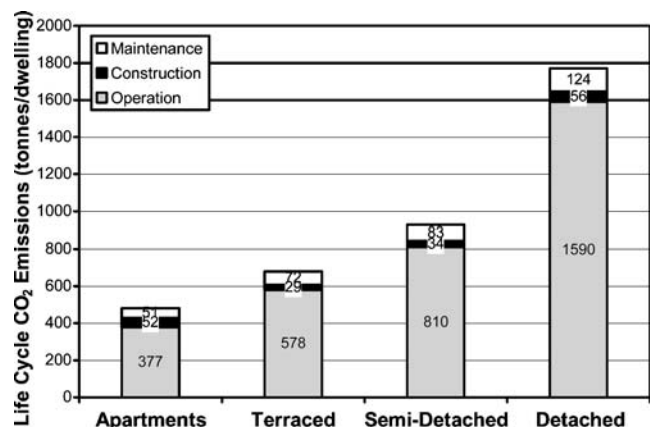


Fig. 10 CO<sub>2</sub> emissions for each dwelling type excluding transport (demolition has been excluded since emissions are negligible compared to other life cycle stages)

583 **4 Conclusions**

584 The life cycle per capita CO<sub>2</sub> emissions in the GDA are  
 585 approximately 50–55% greater in the exurbs and commuter  
 586 towns than in the city centre; corresponding energy  
 587 intensities vary by almost 100%. Per capita residential  
 588 emissions over a 100-year life span taking into account  
 589 construction, operational, transport, maintenance and  
 590 demolition energy requirements were 323, 378, 482 and  
 591 496 tonnes, respectively for the urban, suburban, extra-  
 592 urban and commuter town zones studied.

593 Of the five life cycle stages studied, operational energy  
 594 requirements (predominantly space heating and hot water,  
 595 but including power) contributed most significantly to  
 596 emissions (68%), followed by transport (17%), construction  
 597 (9%) and maintenance/renovation (6%); demolition was  
 598 found to have a negligible effect. Annual per-capita  
 599 operational emissions in Zones 3 and 4 were almost twice  
 600 as great as those in Zone 1 both due to larger dwelling sizes  
 601 and the predominance of detached and semi-detached  
 602 dwellings (with large amounts of exposed walls) in the  
 603 former and the predominance of smaller apartments in the  
 604 latter. The car dominated as the preferred mode of transport  
 605 in all zones; however, this was most pronounced in the  
 606 zones furthest from the city centre (zones 3 and 4) where  
 607 per capita emissions were almost twice those of residents in  
 608 the city centre. It is interesting to note that despite their  
 609 smaller size, the per capita construction energy intensities  
 610 of apartments were approximately one third greater than for  
 611 traditional detached, semi-detached and terraced residences  
 612 due to the greater energy intensity of the structure.  
 613 However, this difference was more than compensated for  
 614 by the significantly lower operational emissions referred to  
 615 above.

616 By 2006 residential development in the GDA completed  
 617 over the 10 years up to and including that year was  
 618 contributing 3,434 kt CO<sub>2</sub>/annum—representing 4.9% of  
 619 national emissions for that year; this comprised 2,108 tonnes  
 620 of recurrent emissions (operational, transport, maintenance)  
 621 and 1,325 kt of once off construction-related emissions.  
 622 Had the development policy required Zone 1-type develop-  
 623 ment and transport modes, then emissions for the year  
 624 2006 would have been 2,892 kt CO<sub>2</sub>—a reduction of  
 625 almost 16% over the actual figure and representing  
 626 approximately 4.1% of national emissions. However, in  
 627 this scenario recurrent emissions would have been reduced  
 628 to 1,417 kt CO<sub>2</sub>—a reduction of 33% over actual levels.

629 **5 Recommendations and perspectives**

630 Irish and other governments' residential sector energy  
 631 policies tend to focus on minimising operational energy

use and emissions. This approach is broadly supported by 632  
 the findings of this paper that approximately two thirds of 633  
 emissions arise from heating and powering dwellings in the 634  
 GDA. However, although operational CO<sub>2</sub> emissions 635  
 dominate those from transport, maintenance and construc- 636  
 tion, assessing only operational emissions would underes- 637  
 timate 100-year emissions by approximately one third. This 638  
 demonstrates the benefits of LCA in evidence-based policy- 639  
 making and highlights the need for policy measures targeted 640  
 at mitigating non-operational emissions. Indeed, the relative 641  
 importance of these emissions will increase as existing 642  
 energy policies further reduce operational emissions. 643

644 An additional weakness of existing policies is the  
 645 emphasis placed on reducing unit area emissions through  
 646 energy efficiency measures. In Ireland, this policy has run  
 647 concurrently with the development of large detached  
 648 dwellings in the area studied. However, the results above  
 649 clearly indicate that large, detached dwellings achieve  
 650 national emissions policy objectives despite having a  
 651 significantly higher environmental impact than smaller  
 652 'attached' dwellings such as apartments. Although other  
 653 social and functional considerations must be taken into  
 654 account, a policy favouring high-density residential devel-  
 655 opment with more modest floor areas would reduce  
 656 emissions. Such a 'high density' planning approach would  
 657 have the added benefit of reducing car dependency since  
 658 residences could be constructed closer to urban centres and  
 659 public transport infrastructure with relatively low emissions  
 660 intensities. The operational and transport emissions from  
 661 the worst performing zones could be reduced by approxi-  
 662 mately 40% with suitable policies.

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